

CHEMISTRY

废弃物中的稀土元素

Rare earth elements from waste

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稀土元素是电子和清洁能源技术中的关键材料。

随着采矿用易得矿物的减少，从废物中回收稀土元素(REE)是电子和清洁能源技术中的关键材料。随着采矿用易得矿物的减少，从废物中回收稀土元素(REE)是电子和清洁能源技术中的关键材料。随着采矿用易得矿物的减少，从废物中回收稀土元素(REE)是电子和清洁能源技术中的关键材料。

Present methods for REE recovery suffer from lengthy purifications, low extractability, and high wastewater streams. Here, we report an ultrafast electrothermal process (~3000 °C, ~1 s) based on flash Joule heating (FJH) for activating wastes to improve REE extractability. FJH thermally degrades or reduces the hard-to-dissolve REE species to components with high thermodynamic solubility, leading to ~2x increase in leachability and high recovery yields using diluted acid (e.g., 0.1 M HCl). The activation strategy is feasible for various wastes including coal fly ash, bauxite residue, and electronic waste. The rapid FJH process is energy-efficient with a low electrical energy consumption of 600 kWh ton⁻¹. The potential for this route to be rapidly scaled is outlined.Present methods for REE recovery suffer from lengthy purifications, low extractability, and high wastewater streams. Here, we report an ultrafast electrothermal process (~3000 °C, ~1 s) based on flash Joule heating (FJH) for activating wastes to improve REE extractability. FJH thermally degrades or reduces the hard-to-dissolve REE species to components with high thermodynamic solubility, leading to ~2x increase in leachability and high recovery yields using diluted acid (e.g., 0.1 M HCl). The activation strategy is feasible for various wastes including coal fly ash, bauxite residue, and electronic waste. The rapid FJH process is energy-efficient with a low electrical energy consumption of 600 kWh ton⁻¹. The potential for this route to be rapidly scaled is outlined.Present methods for REE recovery suffer from lengthy purifications, low extractability, and high wastewater streams. Here, we report an ultrafast electrothermal process (~3000 °C, ~1 s) based on flash Joule heating (FJH) for activating wastes to improve REE extractability. FJH thermally degrades or reduces the hard-to-dissolve REE species to components with high thermodynamic solubility, leading to ~2x increase in leachability and high recovery yields using diluted acid (e.g., 0.1 M HCl). The activation strategy is feasible for various wastes including coal fly ash, bauxite residue, and electronic waste. The rapid FJH process is energy-efficient with a low electrical energy consumption of 600 kWh ton⁻¹. The potential for this route to be rapidly scaled is outlined.Present methods for REE recovery suffer from lengthy purifications, low extractability, and high wastewater streams. Here, we report an ultrafast electrothermal process (~3000 °C, ~1 s) based on flash Joule heating (FJH) for activating wastes to improve REE extractability. FJH thermally degrades or reduces the hard-to-dissolve REE species to components with high thermodynamic solubility, leading to ~2x increase in leachability and high recovery yields using diluted acid (e.g., 0.1 M HCl). The activation strategy is feasible for various wastes including coal fly ash, bauxite residue, and electronic waste. The rapid FJH process is energy-efficient with a low electrical energy consumption of 600 kWh ton⁻¹. The potential for this route to be rapidly scaled is outlined.

INTRODUCTION

稀土元素(REE)是现代电子、清洁能源和汽车工业的战略资源。Rare earth elements (REEs) are strategic resources in modern electronics, clean energy, and automotive industries (1). Concentrated aqueous acid leaching of the REE minerals, followed by biphasic solvent extraction, has been the dominant scheme for REE mass production (1). However, the resource- and pollution-intensive production has a large environmental footprint, where the degradative environmental cost reached \$14.8 billion in 2015, warranting a search for a sustainable solution (2). As the easily accessible REE minerals diminish, the extraction of REE from industrial wastes has gained much attention (3). The applicable secondary wastes include bauxite residue (BR; 以前称为赤泥), 是铝土矿加工用于铝生产的结果; 以及消费电子产品和电动汽车产生的电子废物(电子废物)。

然而, 资源和污染密集型生产的环境足迹很大, 2015年的环境退化成本达到148亿美元, 因此需要寻求可持续的解决方案。

随着易于获取的稀土矿物的减少, 从工业废物中提取稀土元素已受到广泛关注。

适用的二次废物包括粉煤灰(CFA); 铝土矿渣; 以及消费电子产品和电动汽车产生的电子废物(电子废物)。

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归因于难溶REE物种的比例较大, 如REE磷酸盐、锆石和玻璃相。are attributed to the large ratios of hard-to-dissolve REE species such as REE phosphates, zircon, and glass phases (7).

优化酸浸工艺可以通过使用高浓度矿物酸在一定程度上提高萃取率, 例如85 °C至90 °C下15 M HNO₃的萃取率为70%, 85 °C下12 M HCl的萃取率为35%至100%, 具体取决于进料。

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酸浸前CFA的化学或热预处理有助于实现高REE回收率。例如, 通过NaOH水热处理, 然后进行酸浸, 可实现88%的总稀土回收率。

然而, 这些预处理过程通常耗时且能耗高, 这大大降低了利润率。然而, 快速、节能的预处理是从废物中回收稀土元素的必要前提。

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如CFA、BR和电子垃圾所示，活化策略适用于各种二次废物。可行的各种二次废物，如CFA、BR、快速FJH工艺具有可扩展性和高能效，电能消耗低至600千瓦时(kWh)/吨1或12美

元/吨，利润率大于10×。在所有REE中，其中五个(Y, Nd, Eu, Tb和Dy)是最关键的，这取决于它们对清洁能源设备的需求和供应风险。Among all REE, five of them (Y, Nd, Eu, Tb, and Dy) are most critical based on their need for clean energy devices and their supply risk

FJH活化CFA、BR和电子废物中可提取临界REE的百分比比世界上一些最集中的矿石中的百分比高(29)。The percentages of extractable critical REE in the FJH-activated out two to three times. 因此强调了这些当地资源的有用性，这些资源丰富，无需额外开采，被归类为有毒废物，CFA、BR、和 e-waste are two to three times higher than those in some of the most concentrated ores in the world, thus underscoring the usefulness of these local sources that are abundant, requiring no additional mining, and categorized as toxic waste and problematic to stockpile.

RESULTS

CFA中酸可萃取稀土含量

Acid-extractable REE content in CFA

按化学成分分为两种类型的CFA，CFA-F、SiO₂、Al₂O₃和Fe₂O₃的总含量大于70%(wt%)，CFA-C、There are two types of CFA categorized by the chemical composition, CFA-F, with the total content of SiO₂, Al₂O₃, and Fe₂O₃ >70

weight % (wt %), and CFA-C, with a higher abundance of CaO (7). In this work, CFA-F is collected from the Appalachian Basin (App), and CFA-C is collected from the Powder River Basin (PRB) (4), both in the United States. 如X射线衍射所示，CFA由初级非晶相(60-90%)组成，剩余的结晶材料主要包括石英和莫来石相(60 to 90%) (6), and the remaining crystalline materials include mainly quartz and mullite, as shown by the x-ray diffraction

(XRD)图(图1A)。除了Ca在CFA-C中富集外，通过X射线光电子能谱(XPS)(XRD) patterns (Fig. 1A). In addition to the enrichment of Ca in (图1B)和能量色散X射线能谱(EDS)进行的元素分析表明，CFA-F中的C含量较高，这可能是由于煤CFA-C, elemental analyses by x-ray photoelectron spectroscopy (XPS) (Fig. 1B) and energy-dispersive x-ray spectroscopy (fig. S1) show a high C content in CFA-F, which might be caused by the incomplete combustion of coal feeds. 通过热重分析(TGA)，CFA-F中的高C含量在约700 °C下的大量重量损失也很明显(图S2)。The high C content in CFA-F is also evident by the large weight loss at ~700°C by thermal gravimetric analysis (TGA) (fig. S2).

CFA中REE的总量采用氢氟酸(HF): HNO₃消解法(详见材料和方法)。The total quantification of REE in CFA was done by the hydrofluoric acid (HF):HNO₃ digestion method (see details in Materials and Methods) (4). 总REE含量, c_{total}(CFA原始), CFA-F为516 ± 48 mg/kg, CFA-C为418 ± 71 mg/kg(图1C)。The total REE content, c_{total}(CFA Raw), is 516 ± 48 mg kg⁻¹ for CFA-F and 418 ± 71 mg kg⁻¹ for CFA-C (Fig. 1C). App的CFA的REE含量高于PRB 与之前的报告一致。The CFA from App has a higher REE content than that from PRB, consistent with a previous report (4). Acid-leachable REE contents c₀(CFA原料)中的酸浸出REE含量(详见材料和方法)。Acid-leachable REE contents c₀(CFA Raw), were measured by using 1 M HCl or 15 M HNO₃ (see details in Materials and Methods) (4, 9). 对于CFA-F, HNO₃-和HCl可提取的REE含量分别为144 ± 32 mg/kg和160 ± 50 mg/kg (图1C)。For CFA-F, the HNO₃- and HCl-extractable REE contents are 144 ± 32 mg kg⁻¹ and 160 ± 50 mg kg⁻¹ (Fig. 1C), respectively, corresponding to the REE extractability (Y₀) of ~28 and ~31%, respectively. 对于CFA-C, HNO₃-和HCl可提取的REE含量分别为246 ± 71 mg/kg和231 ± 81 mg/kg (图1C)。For CFA-C, the HNO₃- and HCl-extractable REE contents are 246 ± 71 mg kg⁻¹ and 231 ± 81 mg kg⁻¹ (Fig. 1C), respectively, corresponding to the REE extractability of ~59 and ~55%, respectively. 得出的结论是，当酸浓度大于1 M时，对REE浸出能力的影响有限。因此，在以后的实验中，我们使用1 M HCl浸出作为标准方案。It is concluded that the acid concentration has limited effect on the REE leachability once it is >1 M. Hence, in later experiments, we used the 1 M HCl leaching as the standard protocol.

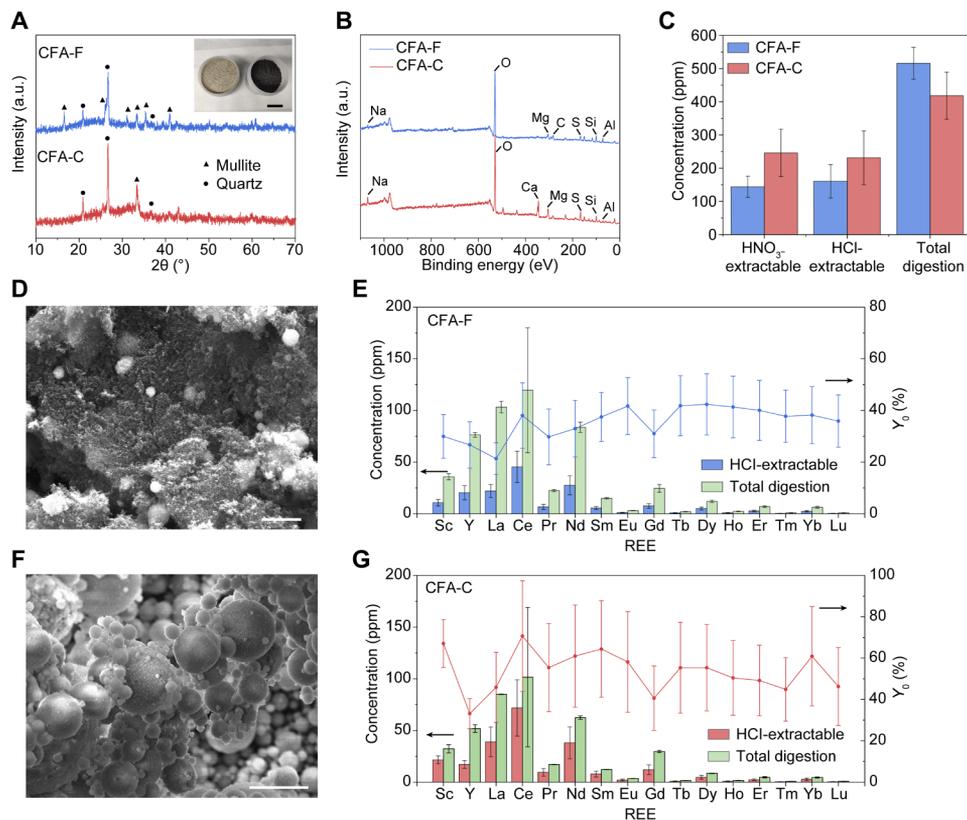


图1. CFA中的酸可萃取REE含量。(A) CFA-F和CFA-C的XRD图谱。插图: CFA-C(左)和CFA-F(右)的图片。比例尺, 4厘米。任意单位。(B) CFA-F和CFA-C的XPS全谱。(C)通过HNO₃浸出(15 M, 85 °C)、HCl浸出(1 M, 85 °C)和总消化, CFA-F和CFA-C中的总REE浓度。(D) CFA-F的SEM图像。比例尺, 2 μm。(E) HCl可萃取REE含量(1 M, 85 °C)和CFA-F中REE的总量以及REE的回收率(Y₀)。(F) CFA-C的SEM图像。比例尺, 5 μm。(G) HCl可萃取REE含量(1 M, 85 °C)和CFA-C中REE的总量以及REE的回收率(Y₀)。(C), (E)和(G)中的所有误差条表示SD, 其中N=3。(C), (E), and (G) represent the SD, where N=3.

CFA-C中稀土元素的酸萃取率高于CFA-F。
 The acid extractability of REE from CFA-C is higher than that from CFA-F. This is consistent with a previous report (7), which attributes the higher extractability to the higher content of easy-to-dissolve REE species like REE oxides in CFA-C. The morphology (SEM) 形态图像如图1D所示, 高碳含量可能会阻碍水酸对含稀土物种的可及性, 导致单个稀土元素的可萃取率较低, 范围为21%-42% (图1E)。 in Fig. 1D, and the high carbon content could retard the accessibility of aqueous acids to REE-bearing species, leading to the low extractability ranging from 21 to 42% for individual REE (Fig. 1E). In contrast, CFA-C is composed of fine, uncovered spheric particles (Fig. 1F), which benefits the acid leaching process, leading to a relatively higher extractability ranging from 33 to 67% for individual REE (Fig. 1G).

**电热活化提高CFA中稀土元素的回收率
 Improved recovery yield of REE from CFA by electrothermal activation**

在FJH的电热活化过程中, CFA原料首先与炭黑(CB)混合, 炭黑作为导电添加剂。 In our electrothermal activation process by FJH, CFA raw materials were first mixed with carbon black (CB), which serves as the conductive additive. The mixture of CFA and CB (~30 wt% CB) was loaded inside a quartz tube between two graphite electrodes (Fig. 2A and Fig. S3). The resistance (R) of the sample was tunable by adjusting the compressive force between the two electrodes that were connected to a capacitance bank of 60 mF (fig. S3). The sample was brought to a high temperature by high-voltage discharging of the capacitors. The detailed experimental parameters are shown in table S1. In a

典型的放电过程, FJH电压为120 V, R为1欧姆, 放电时间为1s, 记录通过样品的电流曲线, 峰值电流约为120 A, 随后电流平台约为7 A (图2B)。 and discharging time of 1 s, the current curve passing through the sample was recorded with the peak current at ~120 A, followed by a current plateau at ~7 A (Fig. 2B). The corresponding real-time temperature curve exhibits a peak temperature up to ~3000°C, followed by the stable heating at ~1150°C (Fig. 2C). The temperature map of the sample during the FJH shows that the temperature distribution is uniform throughout the entire sample without an obvious gradient (fig. S4), demonstrating that the FJH has a homogeneous heating capability. 由于样品的电阻比石墨电极大得多, 放电产生的热量主要施加在样品上 (图S5)。 Because the sample has a much larger resistance than the graphite electrodes, the heat generated by discharging is mainly imposed onto the sample (fig. S5). Thus, the FJH setup has a good durability even though such a high temperature can be achieved. FJH后获得的陶体称为活化CFA c(活化CFA)中的可酸浸出REE含量。 The obtained solid after the FJH is termed as activated CFA (fig. S6). The acid-leachable REE content from the activated CFA, c(activated CFA), was measured by the 1 M HCl leaching procedure. 计算活化CFA(Y)中稀土元素的回收率, 并与CFA原料(Y0)的回收率进行比较(文本S1)。 The recovery yield of REE from the activated CFA (Y) was calculated and compared with that of the CFA raw materials (Y₀) (text S1). 施加一系列50至150 V的FJH电压(图2D)。 A series of FJH voltage ranging from 50 to 150 V were applied (Fig. 2D). 在约120 V下, 活化CFA-F中总REE(1 M HCl, 85 °C)的HCl可浸出含量提高至 329 ± 14 mg/kg(图2D)。 At ~120 V, the HCl-leachable content of total REE (1 M HCl, 85°C) from the activated CFA-F is improved to 329 ± 14 mg kg⁻¹ (Fig. 2D). 这相当于回收率约为64%, 比CFA-F原料(0~31%)提高了约206%。 This corresponds to the recovery yield of Y ~ 64%, representing an increase to ~206% over that of the CFA-F raw materials (Y₀ ~ 31%). 稀土元素的pH依赖浸出动力学。 The pH-dependent leaching dynamics of REE

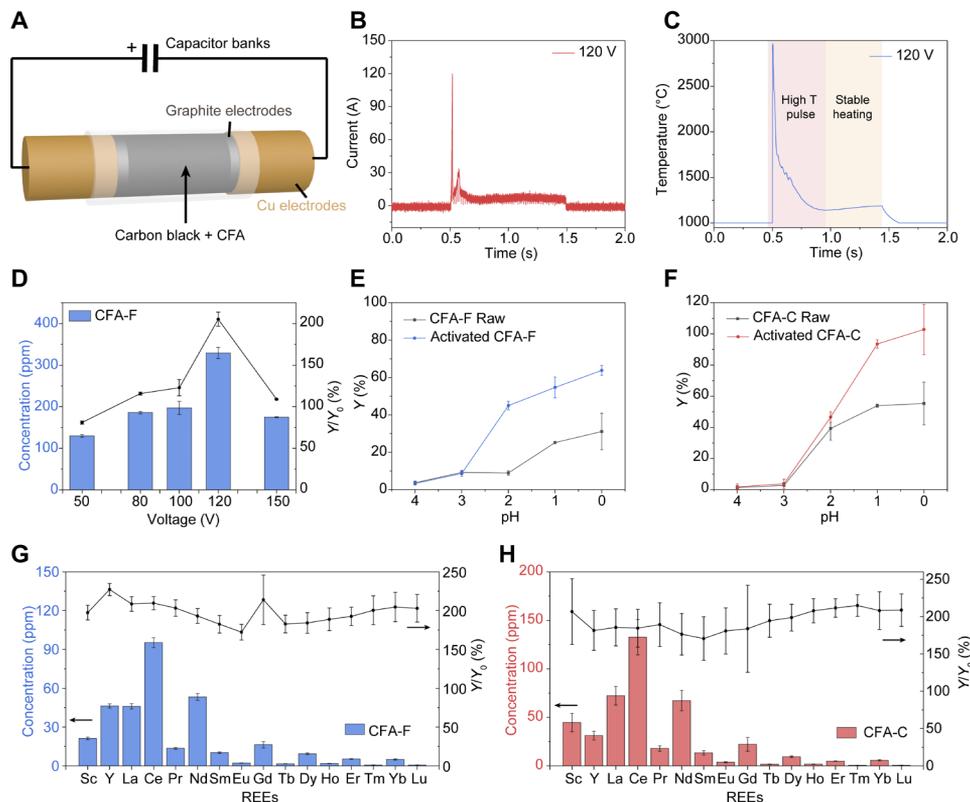


图2. 通过电热活化提高CFA中稀土元素的回收率。 (A) CFA FJH方案。 (B) FJH条件为120 V和1s时的电流曲线。 (C) 在120 V和1s的FJH条件下进行实时温度测量。 (D) 活化CFA-F中HCl可浸出REE含量(1 M, 85 °C), 回收率增加(Y/Y₀)与FJH电压之间的关系。 (E) CFA-F原料和活化CFA-F的pH依赖REE浸出能力。 (F) CFA-C原料和活化CFA-C中REE的pH依赖性浸出能力。 (G) 活化CFA-C中HCl可浸出的单个REE含量(1 M, 85 °C)和回收率的增加。 (H) 活化CFA-F中HCl可浸出的单个REE含量(1 M, 85 °C)和回收率的增加。 Y₀表示从活化CFA原料中回收的稀土回收率, Y表示从活化CFA原料中回收的稀土回收率。 (D)至(H)中的所有误差条表示SD。 All error bars in (D) to (H) represent the SD, where N = 3.

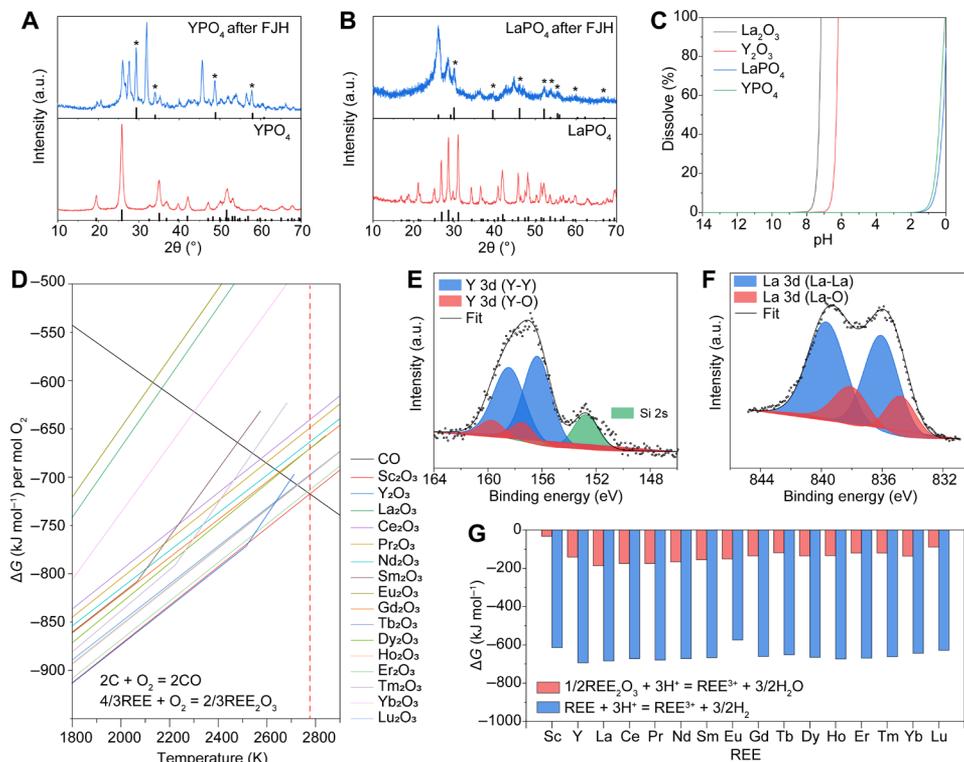


Fig. 3. Mechanism of the improved REE extractability by the electrothermal activation. (A) XRD patterns of YPO₄ after FJH (top) with reference PDF (YPO₄, #11-0254) and YPO₄ after FJH (bottom) with reference PDF (Y₂O₃, #43-0661). (B) XRD patterns of LaPO₄ after FJH (top) with reference PDF (LaPO₄, #35-0731) and LaPO₄ after FJH (bottom) with reference PDF (La₂O₃, #05-0602). (C) Calculated dissolution curves of Y₂O₃, YPO₄, La₂O₃, and LaPO₄ with a mass of 1 g in 100-ml solution. Cl⁻ is used to balance the charge. (D) Ellingham diagram of carbon monoxide and REE oxides. The red dashed line denotes the temperature to reduce Sc₂O₃. (E) XPS fine spectrum of Y₂O₃ after FJH. The Si signal might be from the quartz tube during FJH. (F) XPS fine spectrum of La₂O₃ after FJH. (G) Gibbs free energy change of the REE oxide and REE metal dissolution reactions in acid.

30亿吨已储存在废水池中, 另外 3 billion tons already stored in waste ponds and an additional 150 million tons produced each year, yet just ~3% is currently recycled (41). BR contains a notable amount of REE, for example, a total REE content of ~1000 ppm is found in BR from MYTILINEOS (10). The BR is a dried powder with fine particle size (fig. S11) and has major components including Fe₂O₃, CaCO₃, FeO(OH), and SiO₂ (图4A). BR中的稀土元素通过0.5 M HNO₃的直 接浸出过程提取 (详见材料和方法). The REE in BR was extracted by a direct leaching process using 0.5 M HNO₃ (see details in Materials and Methods) (42). The HNO₃-extractable REE content from BR raw materials is 428 ± 9 mg kg⁻¹ (fig. S12 and Fig. 4B). Similar to CFA, the REE extractability of the BR after the electrothermal activation process is also dependent on the FJH voltage (fig. S12A). At the optimized FJH voltage of 120 V, the HNO₃-extractable REE content increased to 757 ± 30 mg kg⁻¹ (fig. S12B), corresponding to Y/Y₀ ~ 177% of that from the BR raw materials (Fig. 4B). The mechanism of the improvement of REE extractability from BR by the FJH process is presumed to be similar to that of CFA (Fig. 3), because phosphate is also one of the dominant counterions for BR (43). We also applied the FJH strategy for activating e-waste. More than 40 million tons of e-waste are produced globally each year due to the rapid upgrade of personal electronics, with <20% being recycled (44). REEs are widely used in electronics in permanent magnets (14), batteries (13), and capacitors (45). In turn, the recovery of REE from high-grade e-waste has its economic feasibility compared to REE mining from ores. This work in this work is a printed

废弃计算机的电路板 (PCB)。 如图4C所示, 电子垃圾 circuit board (PCB) from a discarded computer. As shown in Fig. 4C, the abundant metals in e-waste include Cu and Al, which are mainly used as the interconnects. The REE in the PCB waste was extracted by 1 M HCl leaching process at 85°C (see details in Materials and Methods). The acid-leachable REE content is 61 ± 4 mg kg⁻¹ from the e-waste raw materials (fig. S13). After the activation process at an optimized voltage of 50 V (fig. S13), the HCl-extractable REE content is increased to 94.6 ± 0.2 mg kg⁻¹, corresponding to Y/Y₀ ~ 156% of that from the e-waste raw materials (Fig. 4D and fig. S13). Different from CFA or BR, the REE species in e-waste are usually in the form of easy-to-dissolve REE metals or oxides (45). However, due to the layered structure, REE通常嵌入基体材料中, 这可能会阻碍湿法冶金过程中的REE提取 (图S14). The FJH process could expose the REE species by cracking the matrices, accelerating the leaching rate and extent of metal extraction (fig. S14).

DISCUSSION
在所有可再生能源中, 根据对清洁能源和供应风险的重要性, 其中五种 (Y, Nd, Eu, Tb和Dy) 被认为是最关键的。 Among all the REEs, five of them (Y, Nd, Eu, Tb, and Dy) are considered most critical based on the importance to clean energy and supply risk (4, 29). 活化CFA-F和CFA-C中HCl可萃取临界REE的百分比分别为34%和26%。 The percentages of HCl-extractable critical REE in activated CFA-F and CFA-C are ~34 and ~26%, respectively (fig. S15, A and B). 提取的临界REE百分比

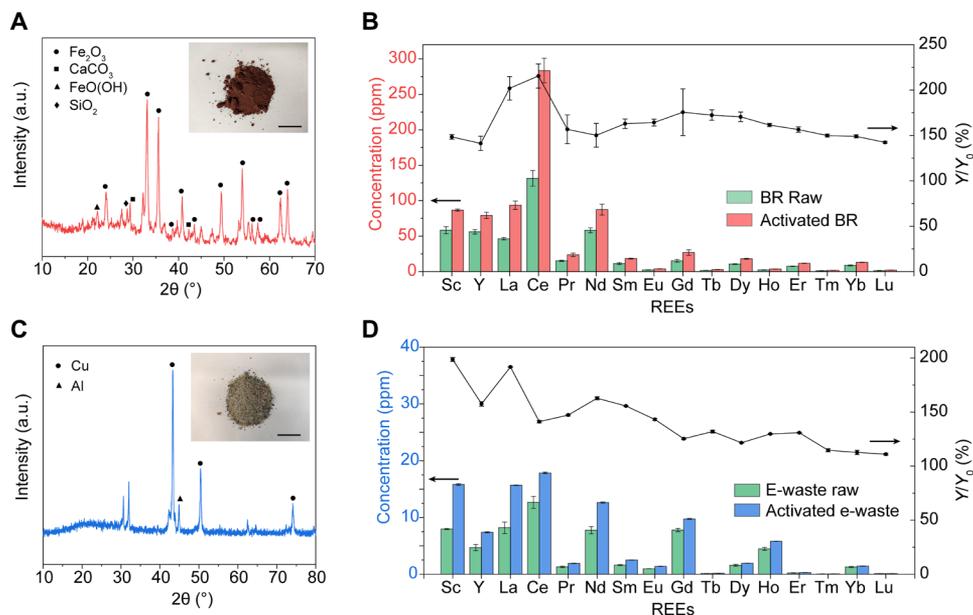


Fig. 4. Recovery of REE from BR and e-waste. (A) XRD pattern of BR. Inset: Picture of BR. Scale bar, 5 μm . (B) BR raw materials and the 120-V FJH-activated BR and the increase in recovery yield (Y/Y_0). (C) XRD pattern of e-waste ground to powder. Inset: Picture of e-waste ground to powder. Scale bar, 5 μm . (D) Acid leachable REE contents (1 M HCl) from e-waste raw materials and the 50-V FJH-activated e-waste and the increase in recovery yield. In (B) and (D), Y_0 represents the REE recovery yield by acid leaching the raw materials, and Y represents the REE recovery yield by acid leaching the activated materials. In (B) and (D), Y_0 represents the REE recovery yield by acid leaching the raw materials, and Y represents the REE recovery yield by acid leaching the activated materials. All error bars in (B) and (D) represent the SD, where $N=3$.

CFA原料和活化CFA几乎相同(图S15, A和B), 表明FJH过程对不同REE没有区别性的活化性能。此外, 活性BR和电子废物中可萃取临界REE的百分比分别约为26%和29%(图S15, C和D)。在废物中可萃取临界REE的百分比远远高于常规矿石中的REE百分比(通常<15%)。例如, 中国白云鄂博的世界最大稀土矿床的临界稀土百分比小于10%。对于像, 拜昂奥博在加拿大具有一个关键的REE百分比为<10% (46)。更高的关键REE在废物中相比传统矿物表示另一个主要优势。

the CFA raw materials and the activated CFA are almost identical (fig. S15, A and B), indicating that the FJH process shows no discriminative activation performance to different REE. In addition, the percentages of extractable critical REE in activated BR and e-waste are ~26 and ~29%, respectively (fig. S15, C and D). The critical REE percentages in these wastes are considerably higher than those in conventional ores (typically <15%) (4). For example, the world's largest REE deposit at Bayan Obo in China has a critical REE percentage of <10% (46). The higher percentage of critical REE in wastes compared to conventional minerals represents another major advantage of the recycling scheme.

我们调查了经济效益, 因为利润率通常是回收利用的支撑因素。由于直接样品加热特性、持续时间短、加热/冷却速度快, FJH工艺是回收再利用的支撑因素。因为直接样品加热特性、持续时间短、加热/冷却速度快, FJH工艺是回收再利用的支撑因素。

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高效节能, 电能消耗量为600 kWh/吨或12美元/吨, 与直接浸出原材料相比, 利润率大于10倍。因为观察到, 在大多数情况下, FJH工艺提高了浸出液中稀土元素和杂质(c(REE)/c(杂质))的含量比(图S18至S20), 表明FJH工艺也有利于随后的REE分离。

Because monazite, (Ce, La, Y, Th)PO₄, and xenotime, YPO₄, are the main commercial sources for REE production (1), the proposed activation strategy could also work for the REE mining to improve the leachability from REE ores. Commercially, alkaline digestion (70% NaOH, 140 °C to 150 °C), and acid digestion (200 °C H₂SO₄) for monazite (49) or acid baking (concentrated H₂SO₄, 200 °C) for xenotime (50). The FJH strategy could be faster and less dependent on the use of concentrated bases and acids. Existing individual elemental separation technologies, such as solvent extraction and ion exchange (51), can be exploited to work with the REE mixtures obtained by FJH because these are often less contaminated than those generated through traditional mining methods (text S4).

MATERIALS AND METHODS

Materials
所用化学品为La(NO₃)₃·6H₂O (99 wt%, Fluka分析), La₂O₃ (99.99%, MilliporeSigma), Y₂O₃ (99.99%, MilliporeSigma), YPO₄ (99.99%, Alfa-Aesar), CB (Cabot, BP-2000), La₂O₃ (99.99%, MilliporeSigma), Y₂O₃ (99.99%, MilliporeSigma), YPO₄ (99.99%, Alfa Aesar), CB (Cabot, BP-2000), H₃PO₄ (85 wt %, Sigma-Aldrich), HCl [37 wt %, ≥99.999% trace metal grade for inductively

耦合等离子体 (ICP) 分析, MilliporeSigma), HNO₃ (67-70 wt%, 用于 ICP 分析的痕量金属等级 coupled plasma (ICP) analysis, MilliporeSigma), HNO₃ (67 to 70 wt %, Fisher Chemical), H₂SO₄ (98 wt%, GR ACS 等级 (保证试剂符合美国化学学会标准), trace metal grade for ICP analysis, Fisher Chemical), H₂SO₄ [98 wt %, MilliporeSigma] 和 HF (48 wt%, 99.99%, 用于 ICP 分析的痕量金属等级, MilliporeSigma). GR ACS grade (guaranteed reagent meeting American Chemical Society standards), MilliporeSigma), and HF (48 wt %, ≥99.99%, LaPO₄ 是通过直接沉淀法合成 (52)). LaPO₄ 是通过直接沉淀法合成 (52). LaPO₄ was synthesized by direct precipitation (52). Concentrated H₃PO₄ was heated to a temperature of 150°C, and La(NO₃)₃·6H₂O was added slowly to allow the emission of NO_x gas. Milky white solid particles precipitated within 30 min, which were then diluted and collected by filtering using a sand core funnel (class F) and diluted to ~20 ml using ultrapure water for ICP-MS measurement. The white solids were dried in an oven (100°C) for 2 hours. The CFA-F samples were collected from App, and CFA-C samples were collected from PRB and provided to our laboratory [see Acknowledgments (4)]. The BR was collected from MYTILINEOS S.A. in Greece and provided to our laboratory (see Acknowledgments). The PCB waste was from a discarded computer. The PCB was cut into small pieces and then ground into powders using a hammer grinder (Dade, DF-15).

FJH system and process

The electrical diagram of the FJH system is shown in fig. S3A. The picture of the FJH system is shown in fig. S3B. The description of each electrical component of the system is listed in the Supplementary Materials. In a typical experiment, the secondary wastes (CFA, BR, or e-waste) were mixed with CB with a mass ratio of 2:1 by using a ball miller (MSE Supplies, PWV1-0.4L). The CB served as the conductive additive. The 200-mg mixture (133 mg of waste and 67 mg of CB) was added into a quartz tube (inner diameter of 8 mm and outer diameter of 12 mm). The resistance was controlled by compressing the two electrodes. The samples were loaded into the jig (fig. S3C), and the electrodes were connected to the capacitor bank. The capacitor bank with a total capacitance of 60 mF was charged by a DC supply. A relay with programmable milliseconds-level relay was used to control the discharging time. The detailed experimental parameters are summarized in table S1. After the FJH, the samples were rapidly cooled to room temperature. For the scaling-up of the process to the gram scale, a larger FJH system with a total capacitance of 0.624 F was built and used (fig. S3D). A larger quartz tube (inner diameter of 16 mm and outer diameter of 20 mm) and a larger reaction jig were used (fig. S3E). Safety caution: There is a risk of electrical shock without proper operation. Safety glasses should be worn to protect the eyes from the bright light emission during the FJH process. The "one hand rule" should be obeyed, and thick rubber gloves extending to the elbows should be used. More implementations were shown in fig. S3.

样品消解、浸出和 ICP-MS/ICP-OES 测量

Sample digestion, leaching, and ICP-MS/ICP-OES measurement
对于 CFA 样品, 进行了酸可萃取 REE 含量测量, 包括 HNO₃ 和 HCl, 以及总 REE 定量。对于 CFA 样品, 进行了酸可萃取 REE 含量测量, 包括 HNO₃ 和 HCl, 以及总 REE 定量。对于 HNO₃ 浸出, 约 50 mg 样品 (CFA 原料) 在 85°C 的 10 ml 浓 HNO₃ (15 M) 中消解 4 小时。对于 HCl 浸出, 约 50 mg 样品 (CFA 原料或 FJH 活化的 CFA) 在约 10 ml HCl (1 M) 中于 85°C 下消解 4 小时。消解后, 使用砂芯漏斗 (F 级) 过滤样品, 并使用超纯水稀释, 以进行 ICP 质谱 (MS) 测量。对于 HCl 浸出, 约 50 mg 样品 (CFA 原料或 FJH 活化的 CFA) 在约 10 ml HCl (1 M) 中于 85°C 下消解 4 小时。消解后, 使用砂芯漏斗 (F 级) 过滤样品, 并使用超纯水稀释至约 20 ml, 以进行 ICP-MS 测量。The pH-dependent leaching dynamics were investigated by using 0.1, 0.01, 0.001, and 0.0001 M

分别以 pH 值为 1 至 4 的 HCl 作为浸出剂。样品 (约 50 mg; CFA 原料或 FJH 活化的 CFA) 在约 10 ml 的 HCl 中于 85°C 下消解 4 小时。CFA raw materials or the activated CFA by FJH were digested and filtered using a sand core funnel (class F) and diluted to ~20 ml using HCl (2 wt %) for ICP-MS measurement. For total REE quantification, CFA raw materials (~30 mg) were digested overnight at 95°C in a mixture of concentrated HF (48 wt %, 2 ml) and concentrated HNO₃ (15 M, 2 ml). The sample was then dried in an oven at 100°C and redigested overnight at 95°C in a mixture of concentrated HNO₃ (15 M, 1 ml), H₂O₂ (30 to 32%, 1 ml), and ultrapure water (5 ml, high-performance liquid chromatography grade). After the digestion, all the solids were dissolved, and the sample was diluted to 50 ml using ultrapure water for ICP-MS measurement. For the BR samples, HNO₃-extractable REE contents were measured. Samples (~50 mg; BR raw materials or the activated BR by FJH) were digested using HNO₃ (0.5 M) at room temperature for 24 hours. The sample was filtered using a sand core funnel (class F) for ICP-MS measurement.

对于电子废物样品, 测量了 HCl 可提取的 REE 含量。对于 e-waste 样品, HCl-可提取的 REE 含量在 85°C 下用 10 ml HCl 样品 (约 50 mg; 电子废物原料或 FJH 活化的电子废物) 在 85°C 下用 10 ml HCl (1 M) 消解 4 小时。然后使用砂芯漏斗过滤样品, 并使用超纯水稀释至 20 ml, 以进行 ICP-MS 测量。The sample was then filtered using a sand core funnel and diluted to 20 ml using ultrapure water for ICP-MS measurement. For the CB as control, HCl-可提取的 REE 含量在 85°C 下用 10 ml HCl 样品 (约 50 mg; 电子废物原料或 FJH 活化的电子废物) 在 85°C 下用 10 ml HCl (1 M) 消解 4 小时。然后使用砂芯漏斗过滤样品, 并使用超纯水稀释至 20 ml, 以进行 ICP-MS 测量。The sample was then filtered using a sand core funnel and diluted to 20 ml using ultrapure water for ICP-MS measurement.

ICP-MS 测量使用 PerkinElmer Nexion 300 ICP-MS 系统进行。在测量中, 主要的 REE 干扰是较轻元素的氧化物。Nexion 300 ICP-MS system. In the measurement, the primary REE interferences were oxides of lighter elements. To avoid the introduction of notable error, the formation of oxides was monitored using a mass ratio of 156:140 (CeO/Ce), which was kept below 3%. The detection limits for REE are in the level of 0.5 to 5 parts per trillion (ppt). The REE mixture standard was used (MilliporeSigma; #10 mg/l; Sc, Y, La, Ce, Pr, Nd, Sm, Eu, Gd, Tb, Dy, Ho, Er, Tm, Yb and Lu in 5 wt% 硝酸中)。periodic table mix 3 for ICP; 16 elements; 10 mg liter⁻¹ each; Sc, Y, La, Ce, Pr, Nd, Sm, Eu, Gd, Tb, Dy, Ho, Er, Tm, Yb, and Lu in 5 wt % nitric acid). In our analyses, the standards were in the concentration range of 1 to 1000 parts per billion. All the analyzed samples were carefully diluted into the concentration range, which is at least two orders of magnitude higher than the detection limit of quantification. All the samples were measured three times to afford the SDs.

还使用 ICP-MS 测量了浸出液中的杂质, 包括铝、硅、铁、镍、铜、钴和铜。The impurities, including Al, Si, Fe, Mg, Zn, Co, Ni, Cr, and Cu, in the leachates were also measured using the ICP-MS. The mixture standard was used (MilliporeSigma, periodic table mix 1 for ICP; 33 elements; 10 mg liter⁻¹ each; Al, As, Ba, Be, Bi, B, Ca, Cd, Cs, Cr, Co, Cu, Gd, Ga, Ge, Hf, In, Fe, Pb, Li, mg, Mn, Ni, P, K, Rb, Se, Si, Ag, Na, Sr, Te, Ti, V, and Zn, in 10 wt % HCl 硝酸溶液中)。Cu, Ga, In, Fe, Pb, Li, Mg, Mn, Ni, P, K, Rb, Se, Si, Ag, Na, Sr, Te, Ti, V, and Zn in 10 wt % nitric acid containing HF traces). ICP-MS 同位素为 m/z (质量/电荷比) = 40, 因此很难使用 ICP-MS 来定义 Ca。因此, ICP 光发射光谱法 (ICP-OES) 被用于量化钙。ICP-OES measurement was conducted using a PerkinElmer Optima 8300 ICP-OES.

表征

使用 FEI Quanta 400 ESEM 场发射扫描电镜在 10 kV 下获得 SEM 图像。SEM images were obtained using a FEI Quanta 400 ESEM field emission microscope at 10 kV. XPS spectra were collected using a PHI Quanterra XPS system under the pressure of 5 × 10⁻⁹ torr.

使用0.1 eV的步长和26 eV的通过能收集元素XPS光谱。Elemental XPS spectra were collected using a step size of 0.1 eV and a pass energy of 26 eV. The XPS spectra were calibrated using the standard C 1s peak at 284.8 eV. XRD patterns were obtained using a Rigaku D/Max Ultima II system with a Cu K α radiation ($\lambda = 1.5406$ Å). TGA was conducted in air at a heating rate of 10°C min⁻¹ to 1000°C using a Q-600 Simultaneous TGA/DSC (differential scanning calorimetry) from TA Instruments. The temperature measurement was conducted using an infrared thermometer (Micro-Epsilon) with a temperature measurement range of 1000° to 3000°C and a time resolution of 1 ms. The optical image of the sample during FJH was captured using an ultrafast camera (Chronos 1.4). The color image was converted into an intensity matrix using MATLAB based on which the temperature map was obtained by fitting the Stefan-Boltzmann law

$$j = \sigma T^4 \quad (1)$$

其中 j 是黑体辐射发射度，是一个比例常数， T 是热力学温度， σ 是Stefan-Boltzmann常数， T 是热力学温度。

计算

溶解曲线根据Visual MINT-EQ 3.1计算。添加额外Cl⁻以补偿费用余额。The dissolution curves are calculated on the basis of Visual MINT-EQ 3.1. Extra Cl⁻ is added to compensate for the charge balance. 由于缺乏稀土氧化物的K_{sp}数据，根据相关的稀土氢氧化物估算稀土氧化物的溶解度[例如，使用Y(OH)₃估算Y₂O₃的溶解]。The solubility of the REE oxide is estimated on the basis of the related REE hydroxide [e.g., Y(OH)₃ was used to estimate the dissolution of Y₂O₃]. 然而，氢氧化物的计算结果与之前报告中氧化物对应的计算结果吻合良好。The calculated results of hydroxides match well with the ones calculated from the oxide counterparts in a previous report (7). 将质量为1g的样品溶解在100 ml溶液中，以获得溶解百分比。The samples with mass of 1 g were dissolved in 100-ml solution to get the dissolution percentage.

SUPPLEMENTARY MATERIALS

Supplementary material for this article is available at <https://science.org/doi/10.1126/sciadv.abm3132>

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